



# Climate change and elevated CO<sub>2</sub> favor forest over savanna under different future scenarios in South Asia

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## Abstract.

South Asian vegetation provides essential ecosystem services to the region and its 1.7 billion inhabitants that are closely linked to its land-use forms and carbon storage potential. Yet, biodiversity is threatened by climate and land-use change. Understanding and assessing how ecosystems respond to simultaneous increases in atmospheric CO<sub>2</sub> and future climate change is of vital importance to avoid undesired ecosystem change. A failure to react to increasing CO<sub>2</sub> and climate change will likely have severe consequences for biodiversity and humankind. Here, we used the aDGVM2 to simulate vegetation dynamics in South Asia under RCP4.5 and RCP8.5, and we explored how the presence or absence of CO<sub>2</sub> fertilization influences vegetation responses to climate change. Simulated vegetation under both RCPs without CO<sub>2</sub> fertilization effects showed decrease in tree dominance and biomass, whereas simulations with CO<sub>2</sub> fertilization showed an increase in biomass, canopy cover, and tree height and a decrease in biome-specific evapotranspiration by the end of the 21st century. The model predicted changes in above ground biomass and canopy cover that trigger biome transition towards tree-dominated systems. We found that savanna regions are at high risk of woody encroachment and transitioning into forest. We also found transitions of deciduous forest to evergreen forest in the mountain regions. C<sub>3</sub> photosynthesis dependent vegetation was not saturated at current CO<sub>2</sub> concentrations and the model simulated a strong CO<sub>2</sub> fertilization effect with the rising CO<sub>2</sub>. Hence, vegetation in the region will likely remain a carbon sink. Projections showed that the bioclimatic envelopes of biomes need adjustments to account for shifts caused by climate change and eCO<sub>2</sub>. The results of our study help to understand the regional climate-vegetation interactions and can support the development of regional strategies to preserve ecosystem services and biodiversity under elevated CO<sub>2</sub> and climate change.

## 20 1 Introduction

Global climate has been identified as the primary determinant of large-scale natural vegetation patterns (Overpeck et al., 1990). Climate change has affected global vegetation pattern in the past and caused numerous shifts in plant species distribution over



the last few decades (Chen et al., 2011; Thuiller et al., 2008). It is expected to have even more pronounced effects in the future that may lead to drastically increasing species extinction rates in various ecosystems (Brodie et al., 2014). Natural ecosystems have been and continue to be exposed to increased climate variability and abrupt changes caused by increased intensity and frequency of extreme events such as heat waves, drought and flooding (Herring et al., 2018). At the same time, they are under severe pressure due to anthropogenic disturbance and land conversion. Rising levels of atmospheric CO<sub>2</sub> are a strong driver of climate-induced vegetation changes (Allen et al., 2014). Anthropogenic CO<sub>2</sub> emissions account for approximately 66% of the total anthropogenic greenhouse forcing (Forster et al., 2007) and are thus largely responsible for contemporary and future global climate change (Parry et al., 2007). Rising CO<sub>2</sub> is expected to alter distributions of plant species and ecosystems (Parry et al., 2007) both indirectly through its influence on global temperatures and precipitation patterns (Cao et al., 2010), two main drivers of vegetation dynamics, and directly via its physiological effects on plants (Nolan et al., 2018). It is therefore of vital importance to understand how ecosystems respond to simultaneous increases in atmospheric CO<sub>2</sub> and temperature, changes in precipitation regime, and altered ecosystem water balance in order to avoid critical ecosystem disruptions and the resulting consequences for biodiversity and humankind.

Increases in temperatures, decreases in precipitation as well as changes in precipitation seasonality can cause loss of vegetation biomass. Elevated atmospheric CO<sub>2</sub> may influence vegetation due to the CO<sub>2</sub> fertilization effect (Curtis and Wang, 1998; Norby and Zak, 2011) that can influence photosynthesis, respiration, decomposition (Doherty et al., 2010), evapotranspiration (ET) and biomass accumulation (Frank et al., 2015). These CO<sub>2</sub> effects on plant growth and competition can alter community structure, ecosystem productivity, climatic niches of ecosystems and biome boundaries (Nolan et al., 2018; Wingfield, 2013)). The physiology of C<sub>3</sub> plants leads to increased carbon fixation efficiency at elevated levels of atmospheric CO<sub>2</sub>, improves their ability for carbon uptake, and thereby increases carbon sequestration (Leakey et al., 2009; Norby and Zak, 2011) as well as plant water use efficiency (Soh et al., 2019).

Long-term Free-Air Carbon dioxide Enrichment (FACE) experiments have demonstrated the productivity-enhancing effect of elevated CO<sub>2</sub> (Norby and Zak, 2011). These effects are due to the fact that C<sub>3</sub> photosynthesis is not saturated at current atmospheric levels and thus C<sub>3</sub> plants benefit from increasing CO<sub>2</sub> (Ainsworth and Rogers, 2007). In contrast, C<sub>4</sub> plants are at their physiological optimum at current atmospheric CO<sub>2</sub>. Increasing CO<sub>2</sub> concentration has been associated with woody cover increase in structurally open tropical biomes such as grasslands and savannas (Stevens et al., 2017). This widespread proliferation of woody plants in arid and semiarid ecosystems has been attributed to increased water use efficiency in C<sub>3</sub> plants that facilitates woody sapling establishment and growth due to higher drought tolerance (Kgope et al., 2010; Stevens et al., 2017). Change in vegetation distribution and altered vegetation structure feed back on climate by altering fluxes of energy, moisture and CO<sub>2</sub> between land and atmosphere (Friedlingstein et al., 2006). Feedback mechanisms also involve vegetation-mediated changes in albedo, evapotranspiration, surface roughness and land-atmosphere fluxes (Field et al., 2007; Richardson et al., 2013).

Ecosystem-level ET is a key ecophysiological process in the soil-vegetation-atmosphere continuum (Feng et al., 2017). Annually, 64% of the total global land-based precipitation is returned to the atmosphere through ET (Zhang et al., 2016). Environmental change and concurrent vegetation changes alter ET and affect water availability (Mao et al., 2015), especially



in arid and semiarid regions. In these regions, ET affects surface and subsurface processes such as cloud development, land surface temperature, and groundwater recharge (Fisher et al., 2011).

60 South Asia is home to approx. 1.7 billion people and is one of the regions most vulnerable to climate change (Eckstein et al., 2018). It hosts four of the world's biodiversity hotspots (Myers et al., 2000) and harbors different biome types ranging from tropical in the south to temperate in the north at the fringe of the Himalayas. These hotspot are characterized by high levels of diversity and endemism, and they are threatened by climate change and anthropogenic pressure such as land-use (Deb et al., 2017). For instance, woody encroachment due to rising CO<sub>2</sub> threatens South Asian savannas (Kumar et al., 2020) and sifting  
65 cultivation in north eastern part of South Asia threatens biodiversity (Bera et al., 2006).

Due to the absence of long-term field experiments such as FACE experiments, in the dominant biome types of the region, modeling studies are valuable tools to close existing knowledge gaps. Dynamic vegetation models (DGVMs, Prentice et al., 2007) are particularly well-suited to address questions that focus on vegetation response to changing environmental drivers, e.g., climate and CO<sub>2</sub>. While most DGVM studies in South Asia focused on vulnerability of forests to climate change  
70 (Chaturvedi et al., 2011; Ravindranath et al., 2006, 1997) they often overlooked the severely threatened savanna biome. They were further limited when using contemporary environmental conditions to pre-define bioclimatic limits of plant functional types (PFTs), and when using fixed eco-physiological parameters, for example to model carbon allocation. Moreover, many DGVMs do not account for life history, eco-evolutionary processes and trait variability among individual plants (Kumar and Scheiter, 2019). While some global-scale studies have investigated the potentially disruptive effect of increasing CO<sub>2</sub> on natural  
75 vegetation, carbon sequestration and biome boundaries (e.g., Hickler et al.; Sato et al., 2007; Smith et al.), detailed modeling studies focusing explicitly on South Asia have not been conducted. The physiological effects of increased CO<sub>2</sub> and climate change on South Asian vegetation is uncertain and needs to be resolved narrowed in order to improve understanding of regional ecosystem functioning as well as implications for biodiversity conservation.

To address the knowledge gaps in existing studies, we used the aDGVM2 (adaptive dynamic global vegetation model version  
80 2), an individual- and trait-based vegetation model that combines elements of traditional DGVMs (Prentice et al., 2007) with a new approach that allows community assembly resulting from environmental filtering applied to traits of modeled plant individuals (Langan et al., 2017; Scheiter et al., 2013). The aDGVM2 thus overcomes the shortcoming of using pre-defined and static PFTs. Originally, aDGVM2 had been tested for Amazonia Langan et al. (2017) and Africa (Gaillard et al.; Pfeiffer et al., 2019). In order to adapt it to South Asian ecosystems and to capture the diversity of South Asian ecosystems, we  
85 included C<sub>3</sub> grasses, improved ecophysiological processes such as the leaf energy budget in order to estimate leaf temperature, implemented separate temperature sensitivities for C<sub>3</sub> and C<sub>4</sub> photosynthetic capacity ( $V_{\text{cmax}}$  and included snow in the water balance model.

In this study we used the updated version of aDGVM2 and addressed the following questions:

1. How do projected changes in climate and CO<sub>2</sub> following two Representative Concentration Pathways (RCP8.5 and  
90 RCP4.5, Meinshausen et al., 2011) change the distribution, boundaries and climatic niches of biomes in South Asia?



2. How does the relationship between projected biomass, ET, temperature and precipitation change in response to CO<sub>2</sub> fertilization?
3. What is the sensitivity of predicted changes in relation to presence and absence of CO<sub>2</sub> fertilization?

Based on our results we analyzed climate-vegetation interactions to improve our understanding on how to manage and mitigate  
95 impacts on biomes under climate change and increasing CO<sub>2</sub>.

## 2 Methods

### 2.1 Description of the study region

Approx. 1.7 billion people populate South Asia, i.e., the Indian subcontinent, Afghanistan and Myanmar. South Asia incorporates a wide range of bio-climatic zones with distinctive biomes, ecosystem types, communities and species (Rodgers and  
100 Panwar, 1988). Climatic conditions are controlled by interactions between the South Asian summer monsoon system and the region's complex topography. The climatic envelope ranges from tropical arid and semi-arid regions in the west, to humid tropical regions supporting rainforests in the northeast and temperate vegetation at the fringe of the Himalaya. Excluding the Himalayan regions, South Asia has a mean annual temperature of approximately 24°C with very low spatial variability. Mean annual precipitation (MAP) is 1190 mm, ranging from less than 500 mm in the warm desert zone in the west to more than 3500  
105 mm in the northeast. The steep elevation gradients ranging from sea level to 8800 m result in a rich diversity of ecosystems that can alternate in areas of a few hundred square-kilometers. Topography is recognized as a strong driver of ecological patterns, for example those related to forest structure and composition, floristic diversity, and soil fertility (Gallardo-Cruz et al., 2009; Jucker et al., 2018; Sinha et al., 2018). South Asia hosts four major global biodiversity hotspots, namely the Western Ghats, Himalayas, Indo-Myanmar and Sri Lanka (www.conservation.org, Conservation International, 2013, Myers et al., 2000). These  
110 hotspots include a wide diversity of ecosystems such as mixed wet evergreen, dry evergreen, deciduous, and montane forests. Further vegetation types are alluvial grasslands and subtropical broadleaf forests along the foothills of the Himalayas, temperate broadleaf forests in the mid hills, mixed conifer and conifer forests in the higher hills, savanna in the Deccan region and southern part of Malaysia, and alpine meadows above the tree line (Conservation International, 2013).

### 2.2 Model Description

115 For this study we used aDGVM2, a DGVM that uses a dynamic trait approach. All details of the model are provided in Scheiter et al. (2013), Langan et al. (2017) and Gaillard et al.. We summarize main features of aDGVM2 in the supplementary material. To adapt the aDGVM2 to the requirements of the study region, we incorporated new sub-routines into the model. We improved the representation of (a) the water balance by including snow, (b) the carboxylation rate, (c) leaf temperature, and (d) we included C<sub>3</sub> grasses (previous model versions only simulate C<sub>4</sub> grasses).

120 (a) Water balance. In aDGVM2, the soil water module is based on the tipping-bucket concept. As the model was originally developed with strong focus on tropical and subtropical forest and savanna regions, the original model version only considered



water input in form of rain (see Langan et al., 2017). In the updated model version, precipitation is assigned as snow when daily mean air temperature drops below 0°C. Snow accumulates on the soil surface or is added to the top of an existing snowpack. The snowpack persists as long as air temperature remains below 0°C. Once temperature rises above 0°C, water from snowmelt is added to the soil water pool and becomes available to plants. This process may improve the water availability for plants at the beginning of spring, for example in the Himalayan region. Snowmelt ( $S_{melt}$ , mm/day) is calculated following (Choudhury et al., 1998) as

$$S_{melt} = 1.5 + K_m P_{precip} (T_a - T_{snow}) S_{pack}, \quad (1)$$

where  $K_m$  is the coefficient of snowmelt (0.007 mm/day/°C),  $S_{pack}$  is the depth of the snowpack,  $T_a$  is daily mean air temperature (°C),  $P_{precip}$  is precipitation (mm/day) and  $T_{snow}$  is the maximum temperature where precipitation falls as snow (0°C). We do not consider insulation effects of the snowpack in the model.

(b) Carboxylation rate.

In earlier versions of aDGVM2, leaf-level photosynthesis was calculated at population level, i.e., it was assumed that all plants of a simulated representative vegetation stand have the same leaf-level photosynthetic rate. Only C<sub>3</sub>- and C<sub>4</sub>-type photosynthesis were distinguished. We therefore implemented new routines to calculate photosynthesis at a daily time step for each individual plant. This version of aDGVM2 now incorporates an empirical relation between specific leaf area ( $A_{SLA}$ , mm<sup>2</sup>/mg) and leaf nitrogen content per unit area ( $N_a$ , g/m<sup>2</sup>) following Sakschewski et al. (2015),

$$N_a = 6.89 A_{SLA}^{-0.571}, \quad (2)$$

The standard maximum carboxylation rate of rubisco per leaf area ( $V_{cmax,25}$ , μmol/m<sup>2</sup>/s) was derived from the TRY database (Kattge and Knorr (2007) by Sakschewski et al. (2015) and is calculated as

$$V_{cmax,25} = 31.62 N_a^{0.501}, \quad (3)$$

where  $V_{cmax,25}$  is  $V_{cmax}$  at 25°C.

In the model,  $A_{SLA}$  is linked to the matric potential at 50% loss of xylem conductance (P50, see Langan et al., 2017). The trade-off between  $A_{SLA}$  and  $V_{cmax}$  mediated by leaf traits ( $N_a$ ) introduces variability in the spectrum of tree growth strategies in aDGVM2. The effect of temperature on photosynthesis is well-described (Kirschbaum, 2004), and temperature may influence photosynthesis both directly, via temperature-dependency of enzyme-mediated metabolic rates of carboxylation and the Calvin cycle (Sharkey et al., 2007), and indirectly via its effect on transpiration and plant water uptake and transport (Urban et al., 2017). The maximum carboxylation rate ( $V_{cmax}$ ) increases with temperature until it reaches an optimum, and decreases again at temperatures above the optimum (Kattge and Knorr, 2007) due to reductions in enzyme activity. Above 30°C the electron transport chain is gradually inhibited, and at temperatures above 40°C the denaturation of Rubisco and associated proteins becomes relevant (Lloyd et al., 2008). The temperature dependency of the carboxylation rate ( $V_{cmax}$ ) is expressed as

$$V_{cmax} = \frac{V_{cmax,25} 2^{0.1(T_{leaf}-25)}}{(1 + e^{0.3(T_{low}-T_{leaf})})(1 + e^{0.3(T_{leaf}-T_{upp})})}, \quad (4)$$



where  $T_{leaf}$  is the leaf temperature in °C (see next paragraph for calculation). The photosynthetic model of Collatz et al. (1992) and Collatz et al. (1992) assumes specific values of  $T_{upp}$  and  $T_{low}$  for  $C_3$  and  $C_4$  plants, respectively (Table S1 and Table S2). This temperature range allows plants to grow most efficiently in their plant-specific climatic niches.

(c) Leaf temperature. We calculate leaf temperature following the leaf-level energy budget concept (Gates, 1968). Leaf-level photosynthesis, activity of leaf enzymes and transpiration depend on leaf temperature ( $T_{leaf}$ , °C), calculated as

$$T_{leaf} = T_{air} + \left( \frac{R_n - \lambda E r_{gb}}{\rho C_P} \right), \quad (5)$$

where  $T_{air}$  is air temperature (°C),  $R_n$  is net radiation absorbed by the leaf (MJ/m<sup>2</sup>/day),  $\lambda$  is latent heat of vaporization (MJ/kg),  $E$  is evapotranspiration (m/day),  $r_{gb}$  is the boundary layer resistance (m/s),  $\rho$  is the air density (kg/m<sup>3</sup>) and  $C_P$  is the specific heat of dry air (MJ/kg/°C). Leaf temperature is used to calculate the temperature dependence of  $V_{cmax}$  used in the photosynthesis model routines in equation (4). Absorbed net radiation ( $R_n$ ),  $r_{gb}$  and  $E$  are model state variables calculated from climate input used in aDGVM2 ( $T_{air}$ , long-wave and short-wave radiation). Latent heat of vaporization ( $\lambda$ ,  $\rho$  and  $C_P$  are assumed as constant parameters in this model version.

(d)  $C_3$  grasses.  $C_3$  grasses were not included in previous aDGVM2 versions (Gaillard et al.; Langan et al., 2017; Pfeiffer et al., 2019; Scheiter et al., 2013). We therefore implemented  $C_3$  grasses, following the approach used for  $C_4$  grasses in previous model versions but adjusted the photosynthetic pathway (see Appendix S2 in Langan et al., 2017).  $C_3$  and  $C_4$  grasses use a different leaf-level photosynthesis model (Farquhar et al., 1980) following the implementations of Collatz et al. (1991, 1992). The optimum temperature ranges for carboxylation for  $C_3$  and  $C_4$  grasses are also different (Table S1). Since  $C_3$  grasses have higher cold tolerance than  $C_4$  grasses (Liu and Osborne, 2008), we implemented frost intolerance for  $C_4$  grasses but not for  $C_3$  grasses. Frost is assumed to damage the tissue of  $C_4$  grasses, and in aDGVM2 we kill 10% of the living leaf biomass of  $C_4$  grasses per frost day independent of frost severity. A representation of  $C_3$  grasses in aDGVM2 is necessary for this study to be able to simulate  $C_3$ -dominated biomes such as the  $C_3$  grasslands in the north of South Asia (Quade et al., 1995).

### 2.3 Climate data

We used GFDL-ESM2M climate data for the period 1950 to 2099 from the Inter-Sectoral Impact Model Inter-comparison Project (ISIMIP2), as historical climate simulated by GFDL-ESM2M showed satisfactory performance over for South Asia (McSweeney and Jones, 2016). The general circulation model (GCM) output was bias-corrected in ISIMIP and downscaled to a spatial resolution of  $0.5^\circ \times 0.5^\circ$  (Warszawski et al., 2014). We used average, maximum and minimum air temperatures, precipitation, surface downwelling shortwave radiation and long-wave radiation, near-surface wind speed, and relative humidity at a daily temporal resolution. We used two representative concentration pathways, namely RCP4.5 and RCP8.5 (Meinshausen et al., 2011). These scenarios assume increases in radiative forcing of 4.5 and 8.5 Wm<sup>2</sup> by 2100 (Van Vuuren et al., 2011) and increases of atmospheric CO<sub>2</sub> concentrations to 560 ppm and 970 ppm by 2100, respectively (Van Vuuren et al., 2011).



## 2.4 Projected changes in temperature and precipitation

Mean annual precipitation (MAP) from GFDL-ESM2M shows decreases for South Asia under RCP4.5 and RCP8.5, except for  
185 a small increase under RCP4.5 during the 2090s (2090-2099, Fig. S1). The Western Ghats and eastern Himalayan are projected  
to become wetter under both RCP4.5 and RCP8.5, whereas the western part of the region is projected to become drier by the  
end of the century under both RCPs (Fig. S2). By the end of the 21st century, mean annual temperature (MAT) of South Asia is  
expected to increase between ca. 1°C and 3.5°C under RCP4.5 and between 1°C and 6°C under RCP8.5, relative to the average  
temperature in the baseline period of 2000–2009 (Fig. S1 and Fig. S2). The western parts of the region and the Himalayan  
190 mountains are projected to experience higher increases in temperature than the rest of the region (Fig. S2).

## 2.5 Soil and elevation data

Soil data was obtained from FAO (<http://www.fao.org/soils-portal>, Nachtergaele et al., 2009) and includes information on soil  
properties and types. The soil properties include parameters required by aDGVM2: volumetric water-holding capacity, soil  
hydraulic conductivity, soil bulk density, soil depth, soil texture, soil carbon content, soil wilting point and field capacity (for  
195 details see Langan et al., 2017). A digital elevation model (DEM) at 90m spatial resolution was obtained from the Shuttle  
Radar Topography Mission (SRTM, <http://srtm.csi.cgiar.org>, Jarvis et al., 2008). It was resampled to a spatial resolution of  
0.5° × 0.5°, to match the spatial resolution of climate data. Elevation values were used to estimate surface pressure. We did not  
use slope and aspect in the model.

## 2.6 Model simulation protocol

200 Using the updated version of aDGVM2, we simulated four different scenarios for South Asia at 0.5° × 0.5° spatial resolution.  
We simulated potential natural vegetation until 2099 using daily climate data for RCP4.5 and RCP8.5. For both scenarios,  
simulations were run with CO<sub>2</sub> increase in line with RCP4.5 (hereafter RCP4.5+eCO<sub>2</sub>) and RCP8.5 (hereafter RCP8.5+eCO<sub>2</sub>)  
and with the same climate data but fixed CO<sub>2</sub> after 2005 at 375 ppm for RCP4.5 (hereafter RCP4.5+fCO<sub>2</sub>) and RCP8.5  
(hereafter RCP8.5+fCO<sub>2</sub>). Fixing the CO<sub>2</sub> concentration after 2005 mimics a situation where CO<sub>2</sub> fertilization would not  
205 occur and vegetation only responds to the climate signal. All simulations were conducted with natural fire as implemented in  
aDGVM2.

To ensure that simulated vegetation had sufficient time to adapt to prevailing environmental conditions, we conducted sim-  
ulations for 650 years, split into a 500 year spin-up phase and a 150 year transient phase. For the spin-up phase, we randomly  
sampled years of the first 30 years of daily climate data (1950 to 1979). For the transient phase, we used the sequence of  
210 daily climate data between 1950 and 2099. Trial simulations showed that a 500 year spin-up period is sufficient to ensure that  
vegetation is in a dynamic equilibrium state with environmental drivers.



## 2.7 Model benchmarking and evaluation

For benchmarking of aDGVM2 simulation results, we used five different remote sensing products: aboveground biomass (Saatchi et al., 2011), tree height (Simard et al., 2011), tree cover (Friedl et al., 2010), MODIS evapotranspiration (Zhang et al., 2010) and natural vegetation type (Ramankutty et al., 2010). All remote sensing data sets were aggregated to a  $0.5^\circ \times 0.5^\circ$  spatial resolution, to match the spatial resolution of model simulations by calculating the mean of all values within each  $0.5^\circ$  grid cell, or using nearest neighbor aggregation in the case of vegetation type ("raster" package in R, Hijmans and van Etten, 2012). We first compared model results and observations assuming that the entire study region is covered by natural vegetation (Figs.1 and 2). Then we repeated the comparisons only for areas with predominantly natural cover, i.e., we masked out areas with more than 50% managed land (Figs. S3 and S4, land cover classes 7 'Cultivated and Managed Vegetation' and 9 'Urban and Built-up' in, Tuanmu and Jetz, 2014). We calculated Normalized Mean Squared Error (NMSE) and coefficient of determination ( $R^2$ ) to quantify agreement between data and simulated variables.

## 2.8 Biome classification

The aDGVM2 simulates state variables such as biomass and canopy cover of individual plants in simulated vegetation stands (1 hectare). We used woody canopy area, abundance of shrubs and trees, and grass biomass to classify the simulated vegetation into biome types (Fig. S5). We used 10-year averages of state variables for the periods 2000-2009, 2050-2059 and 2090-2099 to represent the 2000s, 2050s and 2090s, respectively. Biome classification is often study-specific and to some degree subjective (Torello-Raventos et al., 2013). Here we classified areas with canopy cover below 5% as barren if grass biomass was below 100 kg/ha, and as grassland if grass biomass exceeded 100kg/ha. Grassland was classified as  $C_3$  grassland or  $C_4$  grassland based on predominance of respective biomass proportions. Simulated woody individuals were classified as trees if they had three or less stems and as shrubs if they had four or more stems. If tree canopy cover exceeded shrub canopy cover and was between 5% and 45%, then vegetation was classified as woodland if grass biomass was below 100kg/ha and as savanna if grass biomass was above 100kg/ha (Kumar, 2000). Savanna was further separated into  $C_4$  savanna or  $C_3$  savanna based on the dominant grass biomass. Areas with canopy cover greater than 45% were classified as forests if tree cover exceeded shrub cover, or shrubland if shrub cover exceeded tree cover, irrespective of grass biomass (Fig. S5). Forests were subdivided into evergreen and deciduous forest based on the dominance of canopy area of both tree phenology types. Biomes considered in this study were hence  $C_3$  grassland,  $C_4$  grassland, shrubland, woodland, deciduous forest, evergreen forest,  $C_3$  savanna and  $C_4$  savanna (Fig. S5).

Biomes differ in the amount of precipitation they receive and their temperatures. Whittaker plots describe the boundaries of observed biomes with respect to average temperature and precipitation. We used R-package ('plotbiomes', <https://github.com/valentininelav/plotbiomes> by Valentin Ștefan ) to create Whittaker plot based on Ricklefs (2008) and (Whittaker, 1978)(Figs. 8 and S9). We overlaid the simulated biomes on Whittaker plots to look at climatic niches of biomes under current climate and shifts in climatic niches by the end of this century as a result of climate change and elevated  $CO_2$  under both RCPs (see section 3.6).



## 2.9 Calculation of biome-level evapotranspiration

245 For analyzing evapotranspiration change we calculated the amount of water transpired per unit leaf biomass. Simulated ET and leaf biomass for woody plants, C<sub>3</sub> grass and C<sub>4</sub> grass were summed and scaled to the grid level, taking latitudinal variation of grid cell area into account. As leaf biomass influences transpiration we normalized ET to leaf biomass per grid cell. We then calculated the biome-level ET normalized to biomass as ratio of total annual ET to total leaf biomass for respective biomes:

$$E_{biome} = \frac{\sum_{i=1}^G (E_{grid,i} A_{grid,i})}{\sum_{i=1}^G (B_{grid,i} A_{grid,i})} \quad (6)$$

250 where 1, 2, ..., G represent the grid cells of the study area, E<sub>biome</sub> is biome-level ET (mm/kg/year), A<sub>grid,i</sub> is the area of grid cell i (m<sup>2</sup>, E<sub>grid,i</sub> is total ET of grid cell i (mm/year), B<sub>grid,i</sub> is leaf biomass of grid cell i (kg/m<sup>2</sup>). We calculated the percentage change in E<sub>biome</sub> for respective scenarios between the 2010s and 2050s, and between the 2010s and the 2090s.

## 3 Results

### 3.1 Model performance and contemporary vegetation patterns

255 The aDGVM2 captured contemporary large-scale patterns of biomass, canopy cover, tree height and evapotranspiration. Model results agreed well with remote sensing products used for benchmarking (Fig.1). R<sup>2</sup> was 0.61, 0.45, 0.6 and 0.71, and NMSE was 0.48, 0.78, 0.4 and 1.07 for biomass (Saatchi et al., 2011), tree height (Simard et al., 2011), tree cover (Friedl et al., 2010) and evapotranspiration (Zhang et al., 2010), respectively (Figs.1 and 2). Data-model agreement improved when masking out managed land (Tuanmu and Jetz, 2014). R<sup>2</sup> increased to 0.66, 0.71, 0.67 and 0.80, while NMSE increased to 0.43, 0.30, 0.61  
260 and 1.03 for biomass, tree height, tree cover and evapotranspiration, respectively (Fig. S3 and S4). The model performed well in areas with higher fractional cover of natural vegetation, such as the Himalayas, Western Ghats and the northeast of the region, although the model overestimated biomass and canopy area in the Brahmaputra basin in the northeast of the study region (Fig.1, Kumar et al., 2020).

The model simulated evergreen forests along the Himalayan mountains, southern part of the Western Ghats and Sri Lanka,  
265 whereas deciduous forest was simulated in the northern Western Ghats, central India and southern parts of Myanmar (Fig.3a). Savanna was simulated in southern, northern and western parts of India and some regions of central Myanmar. Shrublands were simulated in the arid regions of Pakistan, the western parts of India and Afghanistan. The aDGVM2 simulated woodland in the west of central India, and grassland in the drier regions (Fig.3a). A large proportion of simulated deciduous forest area is in good agreement with that in maps of potential natural vegetation (PNV, Figs.3b,c). However a large quantity of simulated  
270 savanna area is misrepresented as deciduous forest in the map of PNV (Fig.3b, Kumar et al., 2020).

### 3.2 Projected changes in biome distribution pattern

The aDGVM2 projected changes in canopy cover and above ground biomass in response to climate change and CO<sub>2</sub>, and hence, changes in biome type, predominantly from savanna and grassland to deciduous forest. Simulations showed an increase



in the area covered by evergreen and deciduous forests under elevated  $\text{CO}_2$  ( $\text{eCO}_2$ ) under both RCP4.5 and RCP8.5, in contrast to simulations where  $\text{CO}_2$  was fixed after 2005 ( $\text{fCO}_2$ , (Table.1). Under RCP4.5+ $\text{eCO}_2$ , evergreen and deciduous forest cover increased by 3.1% and 21.2% until the 2050s, and 38.0% and 59.1% until the 2090s, respectively. Under RCP8.5+ $\text{eCO}_2$ , evergreen and deciduous forest increased by 24.8% and 45.4% until the 2050s, and 46.5% and 60.2% until the 2090s, respectively. The model simulated a small increase in forest area for RCP4.5+ $\text{fCO}_2$ , where the area increased by 7.9% and 14.4% for evergreen and deciduous forest until the 2090s, respectively. Evergreen forests were mainly distributed along the Himalayas, Western Ghats and eastern parts of the study region under current conditions (2000s, (Fig.4a), but expanded into the south of peninsular India in future periods (2050s and 2090s) under RCP4.5+ $\text{eCO}_2$ . Deciduous forest cover also increased in future periods in central India and along the Himalayas (Figs.4 and S6).

The extent of  $\text{C}_4$  savanna showed a significant decrease under scenarios with  $\text{eCO}_2$ , although in RCP4.5+ $\text{eCO}_2$ , it showed a temporary increase by 12.1% between the 2010s and the 2050s (Table. 1, Fig.4). Simulated  $\text{C}_4$  savanna area decreased by 14.1% until the 2090s under RCP4.5+ $\text{eCO}_2$ . Under RCP8.5+ $\text{eCO}_2$  the model projected a decrease in  $\text{C}_4$  savanna area of 21.6% and 32.2% until the 2050s and the 2090s, respectively. The area covered by  $\text{C}_4$  savanna increased under both RCPs with  $\text{fCO}_2$  (Table.1).  $\text{C}_4$  savannas were mainly located in the northern plain and peninsular India in the baseline period. However, these areas were replaced by deciduous forests in the northern plain and central India, and by evergreen forests in peninsular India and in the southeast of the region by the 2090s under  $\text{eCO}_2$  scenarios (Figs.4a and S6a). The model simulated an area decrease for woodland, shrubland, grassland and  $\text{C}_3$  savanna by the 2090s under all scenarios (Table. 1, Fig. 4). Simulations showed an increase in barren areas in the western part of the region under all scenarios (Figs. 4 and S6, Table. 1).

### 3.3 Projected changes in biomass at biome level

The aDGVM2 predicted an increase in mean biomass for evergreen and deciduous forest in the  $\text{eCO}_2$  scenarios for both RCPs (Table. 2). Under RCP4.5+ $\text{eCO}_2$ , mean above ground biomass in evergreen and deciduous forest increased by 8.1% and 14.4% by the 2050s and 3.8% and 15.7% by the 2090s, relative to the baseline period. The increase is even higher under RCP8.5+ $\text{eCO}_2$  (Table. 2). The mean biomass of woodland decreased under both RCPs except for the 2050s with  $\text{eCO}_2$  scenarios. The mean biomass of grassland increased under RCP4.5, although it decreased for  $\text{C}_4$  grassland under RCP8.5 for both  $\text{fCO}_2$  and  $\text{eCO}_2$  scenarios. Shrublands in the western part of the study region showed an increase in mean biomass under  $\text{eCO}_2$  scenarios except for the 2050s under both RCPs, and a decrease under  $\text{fCO}_2$  for both RCPs (Table. 2). Our results showed that under RCP4.5 and RCP8.5 biomass decreased in the areas along the Himalayas, as well as in the central, north-eastern and western parts of the study region by the end of the century. Modeled biomass decrease is higher under RCP8.5 in these regions (Figs. 5 and S7). Biomass in the central and south-eastern part of the region is projected to increase under both RCPs with  $\text{eCO}_2$  until the 2050s and 2090s, and to decrease in southern India and in parts of western South Asia (Figs. 5 and S7). We found increased biomass in Afghanistan, western Pakistan, Nepal and southern part of Myanmar, and decreased biomass in the western arid part of the study region under both RCPs for both  $\text{eCO}_2$  and  $\text{fCO}_2$  (Fig.7), though the magnitude of change is considerably different (Figs.5 and S7). There were few small areas in the western part of the study region where the model predicted increased



biomass only under  $f\text{CO}_2$  for both RCPs (Figs.7). We found increased biomass over large part of the study region under  $e\text{CO}_2$  for both RCPs shows areas (Figs.7) where  $\text{CO}_2$  fertilization compensates climate change induced biomass die-back.

### 3.4 Projected changes in evapotranspiration at biome level

310 The response of simulated  $E_{\text{biome}}$  varies in different biomes under both RCP4.5 and RCP8.5 (Table. 3). Under the RCP4.5+ $f\text{CO}_2$  scenario the model predicted a decrease in ET in all biomes except for deciduous forest and shrubland where it increased by 1% and 2.1% until the 2050s, and by 0.3% and 11.9% by the 2090s, respectively. Simulated  $E_{\text{biome}}$  under RCP8.5+ $f\text{CO}_2$  for deciduous forest and shrubland increased by 4.2% and 5.2% until the 2050s, and by 5.2% and 16.4% until the 2090s, respectively. The model also predicted increased  $E_{\text{biome}}$  for  $\text{C}_4$  grassland, evergreen forest and  $\text{C}_4$  savanna until the 2090s  
315 under RCP8.5+ $f\text{CO}_2$  (Table.3). Comparisons of the RCP4.5+ $f\text{CO}_2$  and RCP8.5+ $f\text{CO}_2$  scenarios indicated that the former had a higher  $E_{\text{biome}}$  than the latter scenario across all biomes because precipitation decrease is higher in the RCP8.5 scenario than in the RCP4.5 scenario. Under both RCPs with  $e\text{CO}_2$ , the model predicted a decrease in  $E_{\text{biome}}$  across all biomes, except a marginal increased  $E_{\text{biome}}$  for shrubland under RCP4.5 and deciduous forest under RCP8.5 until the 2050s and the 2090s (Table. 3). In general, RCP scenarios with  $e\text{CO}_2$  predicted reduced  $E_{\text{biome}}$  across most of the biomes compared to simulations  
320 with  $f\text{CO}_2$ .

### 3.5 Response of mean ET and mean above ground biomass to climate change

The model predicted a larger increase in absolute annual mean ET (mm/year) under  $e\text{CO}_2$  than  $f\text{CO}_2$  for both RCP scenarios due to the corresponding increase in biomass (Figs. 5 and S7). We compared the spatially averaged annual values of simulated absolute ET with MAP over the period from 1951 to 2099 and found a statically significant relation ( $p$ -value  $< 0.005$ ). We  
325 found that absolute ET is positively correlated with MAP under all four scenarios (Figs.6a and S8a), but weakly correlated with MAT (Figs. 6b and S8b). For a given MAP, the spatially averaged annual value of above ground biomass (AGBM) was lower in scenarios with  $f\text{CO}_2$  than scenarios with  $e\text{CO}_2$  under both RCPs (Figs. 6c and S8c). Spatially averaged annual value of AGBM decreased beyond a MAT of  $23.5^\circ\text{C}$  for both RCPs with  $f\text{CO}_2$ , whereas it increased beyond  $23.5^\circ\text{C}$  under both RCP scenarios with  $e\text{CO}_2$  (Figs.6d and S8d).

### 330 3.6 Impact of climate change on climatic niches of biomes

The climate niches of simulated biomes broadly overlapped with biome niches in the Whittaker scheme (Figs. 8 and S9, Ricklefs, 2008; based on Whittaker, 1975). Under RCP4.5+ $e\text{CO}_2$  and RCP8.5+ $e\text{CO}_2$ , the aDGVM2 simulated shifts of climatic niches of biomes. Evergreen and deciduous forest biomes were predicted to invade the niche space of savannas under  $e\text{CO}_2$  scenarios (Figs.5 and S9). Savannas in turn were predicted to expand their climatic niche to  $\text{MAT} > 30^\circ\text{C}$  by 2099, a climatic  
335 space that was essentially not occupied by current biomes. Under RCP8.5+ $e\text{CO}_2$ , forests completely occupied the climate space currently occupied by savanna (Fig. S9).



In both scenarios with  $fCO_2$ , savanna occupied the climate with  $MAT > 25^\circ C$  and  $MAP$  between 500mm and 1500mm and did not show major woody encroachment. The model predicted that savanna expansion in climate space was higher under RCP8.5+ $fCO_2$  than under RCP4.5+ $fCO_2$  (Figs.8 and S9). Other biomes also experienced shifts in their climate space (Fig. 8). Grasslands and shrublands occupied the region with low  $MAP$  ( $< 500mm$ ), however the results showed that woodland also occupied low  $MAP$  ( $< 800mm$ ) regions, which corresponds to the western arid and semi-arid region of the study region under scenario with  $eCO_2$  (Fig. 8).

## 4 Discussion

### 4.1 Impact of climate change and elevated $CO_2$ on biomes and biomass

Our simulations for RCP4.5+ $eCO_2$  and RCP8.5+ $eCO_2$  showed a strong positive impact on vegetation growth, i.e., increases of biomass, canopy cover and canopy height. Mean biomass in most biomes was projected to increase, but the magnitude of increase differed considerably between different scenarios (Table.2). Projected change in canopy cover resulted in biome transitions. Under future conditions, the spatial distribution, extent and biomass of evergreen forests mostly remained at the current state, and evergreen forests were more resistant to climate change than deciduous forests. Expansion of deciduous forest into open biomes due to increasing woody cover resulted in significant loss of savanna area in the Deccan region under both RCPs with  $eCO_2$  by the end of the century. Transition from deciduous forests to evergreen forest was simulated for the mountain regions of South Asia.

The increasing woody biomass and canopy cover, i.e., woody encroachment, agree with the reported greening trend in South Asia during the last three decades (Wang et al., 2017). Increasing  $CO_2$  has been identified as the most likely driver of increasing biomass in many ecosystems, in studies based on both field observation e.g., FACE experiments, and satellite data (Brienen et al., 2015; Fischlin et al., 2007; Piao et al., 2006; Schimel et al., 2015). Elevated  $CO_2$  affects plants by increasing their photosynthetic rate, growth rate and water use efficiency, leading to an increase in biomass (Leakey et al., 2009; Norby and Zak, 2011). Increased photosynthetic rates under elevated  $CO_2$  are due to an increase in the rate of rubisco carboxylation for  $C_3$  plants, with a concurrent decrease of photorespiratory losses of carbon (Long et al., 2004). Due to the improved carboxylation efficiency,  $C_3$  plants can respond by reducing stomatal conductance, thereby reducing transpirational losses, improving leaf water status, water use efficiency, and favoring leaf area growth (Long et al., 2004; Norby and Zak, 2011). However, it is still unclear how these responses scale to the ecosystem level (Hickler et al., 2015), and how nutrient limitation from the soil may influence ecosystem responses to  $eCO_2$ . Körner (2015) argued that carbon from atmosphere can only be converted into biomass if other factors such as nutrients, temperature and water are not limiting. In the long run, whether ecosystems act as carbon source or sink can be estimated using models that consider all factors that are relevant in the carbon cycle (Fatichi et al., 2014; Körner, 2015). According to our simulations we can conclude that natural vegetation of South Asia likely will remain a carbon sink in the future.



#### 4.2 Impact of climate change and fixed CO<sub>2</sub> on biomes and biomass

Under both fCO<sub>2</sub> scenarios, the spatial distribution of savanna areas remained in its contemporary state. Central India and the  
370 Deccan Plateau showed a transition of deciduous forest to savanna, because forest canopy opened up due to tree mortality  
caused by increasing temperature and reduced MAP. This indicates that plants experience temperature and drought stress under  
fixed CO<sub>2</sub> concentration. These stresses were compensated by CO<sub>2</sub> fertilization in eCO<sub>2</sub> scenarios where the aDGVM2  
simulated increased biomass and woody encroachment in areas affected by climate-induced die-back in fCO<sub>2</sub> simulations.  
This aDGVM2 behavior agrees with, Lapola et al. (2009) who modeled biome shifts from forest to savanna in absence of  
375 CO<sub>2</sub> fertilization for the Amazon region. Changes in precipitation regimes are likely to have a strong influence particularly in  
arid and semi-arid regions, such as grasslands (Verstraete et al., 2009). The complex interactions of inter-annual precipitation  
variability and precipitation seasonality can result in rapid ecosystem transitions (e.g., between alternative stable states with  
high and low vegetation biomass; Holmgren and Scheffer, 2001). The decrease in simulated AGBM after MAT increases beyond  
23.5°C under fCO<sub>2</sub> scenarios can be explained by the longer exposure of vegetation to temperatures beyond the optimum  
380 temperature range of C<sub>3</sub> photosynthesis during the main growing season. This effect was further enhanced by decreasing MAP  
and the absence of CO<sub>2</sub> fertilization. This implies that the increase in MAT above 23.5°C together with weak CO<sub>2</sub> fertilization  
would have negative consequences for carbon sequestration. It also implies that we need a better understanding of impacts  
of heat stress on vegetation and how it interacts with drought and CO<sub>2</sub> fertilization. It is also unclear to what degree thermal  
acclimation may counteract some of the negative effects on plant growth caused by higher temperatures (Lombardozzi et al.,  
385 2015).

#### 4.3 Impact of climate change and CO<sub>2</sub> change on climatic niches of biomes

Elevated CO<sub>2</sub> has a major impact on the climatic niche space of biomes, especially the climatic envelope of savannas. Our  
simulations indicated that forest invaded the niche space currently occupied by savanna by the end of the century. The expansion  
of forests to drier areas corresponds to a widening of their climate niche space under eCO<sub>2</sub>. This expansion is corroborated by  
390 the fact that in absence of CO<sub>2</sub> fertilization the climatic niche of biomes is stable, i.e., biomes occupy the same niche under  
current and future conditions. These findings imply that the bioclimatic boundaries used to define biome niche space are not  
static, but are specific for given CO<sub>2</sub> levels. Therefore, the thresholds of the Whittaker's biomes need to be redefined for a  
high-CO<sub>2</sub> world such that they encompass the altered climatic envelopes of biomes under elevated CO<sub>2</sub> in future (Fig. 5). The  
shift in niche space can be attributed to the shift in plant communities caused by the combined effect of climate change and  
395 elevated CO<sub>2</sub>, which increases plant water efficiency allowing them to grow in harsh condition. These shift can also lead to  
change the characteristics of biomes by altering community structure and ecosystem functions (Chapin et al., 1997).

#### 4.4 Effect of CO<sub>2</sub> on ET and its interaction with climate change

Climate change has direct effects on hydrological processes (Liu et al., 2008). ET and water deficit influence plant productivity  
and distribution (Stephenson, 1998). Higher biomass resulted in increased absolute amounts of ET for eCO<sub>2</sub> scenarios in some



400 parts of the study region under both RCPs by the end 21st century (Figs. 5 and S7). This change can be attributed to the fact  
that plants accumulated higher biomass while reducing stomatal conductance due improved water use efficiency under eCO<sub>2</sub>  
scenarios and resulted in reduced ET per unit leaf biomass (Warren et al., 2011). Our results showed that the strength of the  
CO<sub>2</sub> fertilization effect is relevant when attempting to determine E<sub>biome</sub> at biome level during the 21st century. Biome-specific  
ET decrease was less pronounced under RCP4.5 due to a lower concentration of atmospheric CO<sub>2</sub> compared to RCP8.5. Our  
405 simulated decrease in ET in response to climate change and increasing CO<sub>2</sub> concentration agrees with Kergoat et al. (2002) who  
have reported decreased ET under elevated CO<sub>2</sub> concentration in a chamber experiment. However, reduced ET under eCO<sub>2</sub>  
can reduce regional-scale atmospheric humidity and thereby enhance the vapor pressure deficit (VPD), a driving force for  
water loss, between leaves and the atmosphere, which may partially counteract CO<sub>2</sub>-induced reduction of ET due to decreased  
stomatal conductance. Projected increases in air temperature can also enhance atmospheric water storage capacity and VPD,  
410 and thereby evapotranspiration may increase (Warren et al., 2011). As future climate projections vary spatially and temporally,  
there was high model uncertainty on how ET will respond to changes in precipitation and temperature.

#### 4.5 Implication of woody encroachment for water resources

Water is a limiting factor for vegetation in arid and semi-arid regions. Improved WUE of C<sub>3</sub>-plants under eCO<sub>2</sub> boost produc-  
tivity and therefore facilitated woody encroachment. However, increased woody cover can negatively affect water resources.  
415 Acharya et al. (2018) showed that increased woody cover hinders the downward movement of water in the soil profile as result  
of increased rooting depth and density in more densely-packed woody vegetation. Increased water inception through roots in  
the soil can have negative effects for ground water recharge. It is therefore necessary to control the abundance of woody plants  
in semi-arid regions to control streamflow and enhance groundwater recharge (Bednarz et al., 2001).

#### 4.6 Implication of the projected change for conservation

420 Changes in biome types imply changes in biodiversity, ecosystem function and productivity. Each biome is characterized  
by a range of distinctive ecological processes and functions. The distribution pattern of plants in the mountains is largely  
regulated by the altitude and climatic factor (Saikia et al., 2017). These fragile ecosystems such as forests, in the mountains  
have high species richness and needed to be protected from the ever-increasing anthropogenic pressure and climate change. In  
the other hand, open biomes such as grassland and savanna support high biodiversity (Parr et al., 2014). Pronounced increases  
425 in tree density in grasslands and savannas will alter vegetation structure and reduce grassland biodiversity. Such changes  
will negatively affect savanna-specific ecosystem services such as grazing potential, tourism and wildlife habitat availability  
(Parr et al., 2012). The threat posed to the biodiversity of Asian savannas by climate change is aggravated by inadequate  
management policies that misinterpret them as degraded forest (Ratnam et al., 2016). In this context, management policies  
aim to afforest open biomes, although paleo-ecological evidence indicates that these open biomes are not degraded forest but  
430 ancient ecosystems (Kumar et al., 2020; Ratnam et al., 2016).

In South Asia, biodiversity hotspots have a very unique topography, where climate varies strongly over short distances. As  
global biodiversity hotspots, mountain forest ecosystems in the Western Ghats, the Himalayas and northeastern part of the



study area (Indo-Myanmar) are particularly vulnerable to climate change (Myers et al., 2000) and need targeted management action to mitigate adverse effects. Conservation of these hotspots requires consideration of many different attributes of plant communities, ecosystems, landscapes, and plant diversity, how they will change, and how their ecosystem services are valued.

Conservation methods and policies that can accommodate minimal losses of ecosystem services and provide robust strategies for mitigating climate change impacts should be developed and implemented. In this context, DGVMs facilitate the exploration of vegetation-climate interactions by providing detailed results for different management and climate scenarios. Such an exploration of different possible scenarios is necessary to develop optimized mitigation and conservation strategies for protected areas in biodiversity hotspots. The value of DGVM modelling results lies in their potential to provide insights into multiple future trajectories. Based on the most likely trajectories, the results can be used to tailor best-practice strategies for decision makers that need to manage conservation areas or protected areas Boulangeat et al. (2012).

#### 4.7 Limitations of this modelling study

Our simulation results are constrained by the model formulation and the assumptions underlying aDGVM2. Disagreement between model results and data used for benchmarking can be attributed to the fact that the aDGVM2 simulates potential natural vegetation whereas remote sensing products integrate land use. This implies that enhancing the model to simulate observed land cover patterns would require additional information on anthropogenic impacts. Anthropogenic activities such as deforestation, habitat conversion and urbanization can modify the interactions between climate, plant communities and biomes (Hansen et al., 2001).

In addition data-model disagreement can be explained by uncertainties in aDGVM2 and processes currently not considered. For instance, the aDGVM2 uses carbon allocation parameters that are not easily measurable in the field, which limits the evaluation of simulated mechanisms. The model currently lacks a representation of carbon that plants actively invest into nutrient acquisition (e.g. mycorrhiza) or to build defenses against predation and pathogens (Zemunik et al., 2015) and there is insufficient ecophysiological data from the region, required for parameterization of trait ranges used to simulate plant communities (Kumar and Scheiter, 2019). The complexity of the interactions between global change and biomes as well as biodiversity is difficult to model in absence of such data. Further, the strength of CO<sub>2</sub> fertilization modelled in aDGVM2 may be overestimated, given that the model does not include nutrient limitation (Körner et al., 2005; Terrer et al., 2018). Despite these caveats, we are nonetheless confident to capture general patterns of future global change and its consequences for biodiversity in South Asia.

## 5 Conclusions

We investigated the impact of eCO<sub>2</sub> and climate change on South Asian vegetation with the aDGVM2. The model reproduced the contemporary distribution of biomes, biomass, ET and tree height. In comparison to fCO<sub>2</sub> simulations, we found that climate change and CO<sub>2</sub> fertilization in combination are substantial drivers of biome change, and that elevated CO<sub>2</sub> concentrations also altered the climatic envelope of biomes in addition to causing increases in biomass, tree height and canopy cover.



465 Continued biomass increase indicates that South Asia's natural vegetation will likely remain a carbon sink in the 21st century. Our results also imply woody encroachment posing threat to open biomes and causing transition of savanna biomes to deciduous forest in the future which are particularly critical aspect in the context of biodiversity conservation. Thresholds that define bioclimatic envelopes of biomes are not static, but are specific for given CO<sub>2</sub> concentrations which imply that these thresholds need to be adjusted to account for changed climatic niches caused by eCO<sub>2</sub> in the future. We also found that simulated decrease  
470 in biomass-specific ET is more pronounced in scenarios with eCO<sub>2</sub> than in scenarios with fCO<sub>2</sub> which indicates that water use efficiency will likely increase due to CO<sub>2</sub> fertilization.

The biome transitions simulated under eCO<sub>2</sub> and changing climate indicate the need to adjust ecosystem management, mitigation strategies, and conservation policies for protected areas to allow targeted long-term management. To understand the significance of ecological responses to climate change, it is essential to improve and expand biological monitoring activities  
475 (Loreau et al., 2001). To achieve this, most vulnerable biomes identified by modeling results such as the ones we present in this study could be proposed as high-priority targets for programs that monitor vegetation-climate interactions, productivity and biodiversity Proença et al. (2017).

*Code availability.* The aDGVM2 code as well as scripts to conduct the model experiments and analyze the results are available upon request. Please contact any of the authors.

480 *Author contributions.* DK and SS conceived the study, DK included changes to aDGVM2 for the study region, conducted model simulations, analyzed results, created figures and lead the writing. All authors contributed to the development of aDGVM2 and comments to the text.

*Competing interests.* The authors declare that they have no conflict of interest.

*Disclaimer.* Key words: aDGVM2, climate change, CO<sub>2</sub>-fertilization, biome shift, woody encroachment, evapotranspiration, biodiversity conservation, South Asia

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**Table 1.** Biome cover (in %) for the 2000s, 2050s and 2090s, and percent (%) change in biome cover from the 2000s to 2050s and the 2000s to 2090s under RCP4.5 and RCP8.5 with fixed and elevated CO<sub>2</sub>.  $\Delta$  indicates percentual biomass changes between time periods.

RCP Scenarios	Year	Barren	C <sub>4</sub> Grassland	C <sub>3</sub> Grassland	Woodland	Evg. forest	Dec. forest	Shrubland	C <sub>4</sub> Savanna	C <sub>3</sub> Savanna
RCP4.5 + fCO <sub>2</sub>	2010s	5.6	15.4	4.6	18.2	11.7	17.6	6.9	17.7	2.4
	2050s	6.3	14.8	3.2	15.7	11.2	18.6	6.7	22.1	1.4
	2090s	10.4	12.3	2.3	10.0	12.7	20.1	6.0	24.7	1.4
	$\Delta$ 2050s-2010s	13.0	-3.7	-32.2	-13.6	-4.0	5.6	-3.0	25.4	-39.1
	$\Delta$ 2090s-2010s	87.0	-20.1	-50.0	-45.2	7.9	14.4	-12.7	40.1	-41.3
RCP4.5+ eCO <sub>2</sub>	2010s	5.7	15.2	4.8	18.6	11.5	17.5	6.8	17.5	2.4
	2050s	6.5	13.9	3.5	15.0	11.9	21.2	7.0	19.6	1.3
	2090s	10.4	10.4	2.5	10.7	15.9	27.9	6.2	15.1	0.9
	$\Delta$ 2050s-2010s	13.5	-8.2	-26.9	-19.7	3.1	21.2	3.8	12.1	-44.7
	$\Delta$ 2090s-2010s	82.0	-31.6	-48.4	-42.4	38.0	59.1	-8.4	-14.1	-63.8
RCP8.5 + fCO <sub>2</sub>	2010s	6.3	14.7	4.5	18.8	11.7	17.3	6.3	18.0	2.4
	2050s	8.8	12.3	2.5	14.7	12.9	21.7	6.6	19.0	1.5
	2090s	9.4	15.0	2.5	11.0	10.8	14.2	6.7	29.0	1.4
	$\Delta$ 2050s-2010s	39.0	-16.5	-43.7	-21.9	10.1	25.0	4.1	5.4	-39.1
	$\Delta$ 2090s-2010s	48.8	1.8	-43.7	-41.6	-7.9	-17.9	5.7	61.0	-41.3
RCP8.5 + eCO <sub>2</sub>	2010s	5.9	14.8	4.7	18.0	11.6	17.5	7.1	17.9	2.5
	2050s	9.7	10.5	3.2	13.9	14.5	25.4	7.1	14.1	1.6
	2090s	6.3	11.5	4.2	12.6	17.0	28.0	7.0	12.2	1.3
	$\Delta$ 2050s-2010s	64.9	-29.5	-32.6	-22.9	24.8	45.4	0.7	-21.6	-35.4
	$\Delta$ 2090s-2010s	7.0	-22.2	-10.9	-30.3	46.5	60.2	-1.5	-32.2	-47.9



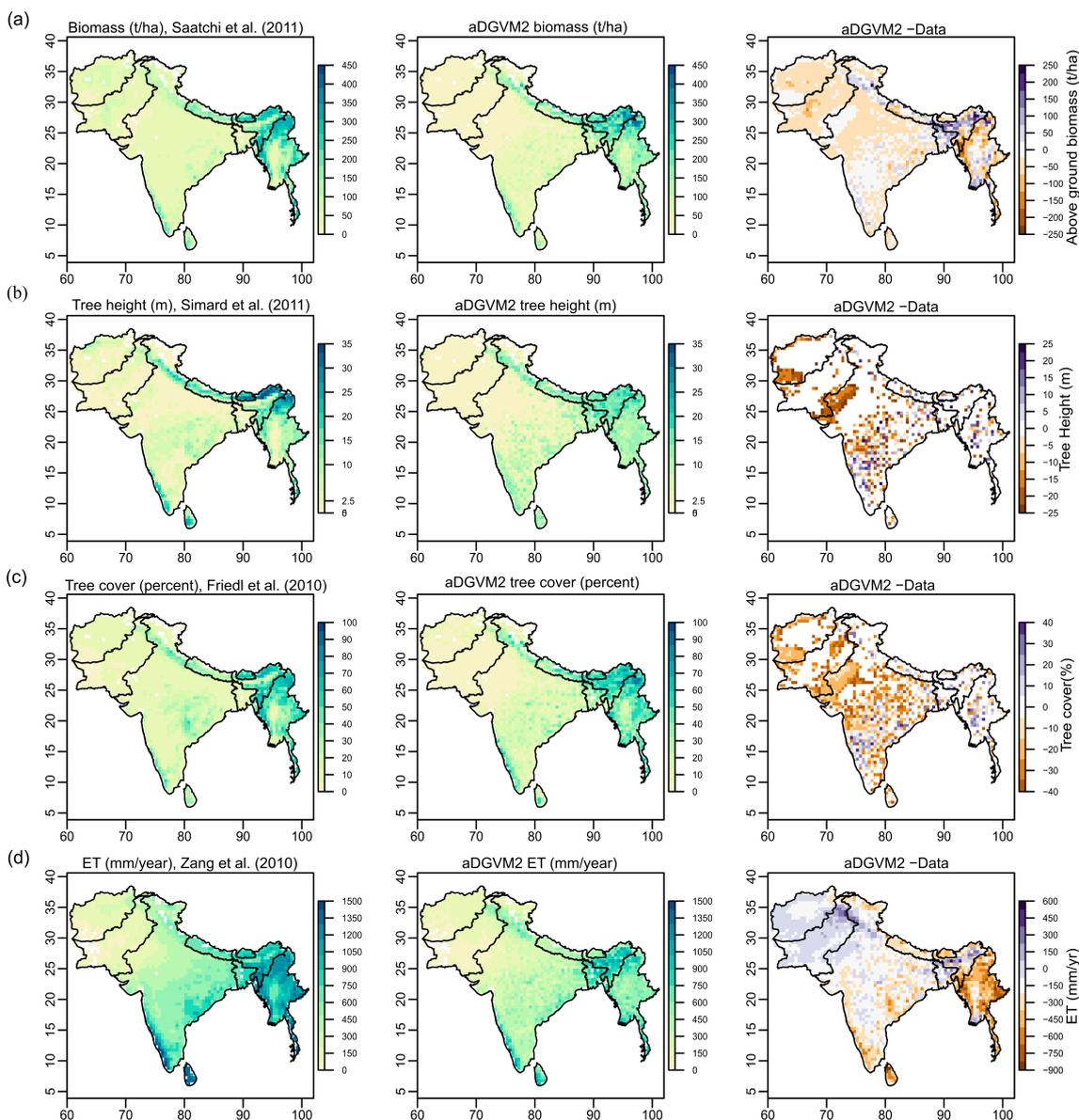
**Table 2.** Mean biomass (in t/ha) within biomes for the 2000s, 2050s and 2090s, and percent (%) change in biomass from the 2000s to 2050s and the 2000s to 2090s under RCP4.5 and RCP8.5 with fixed and elevated CO<sub>2</sub>.  $\Delta$  indicates percentual biomass changes between time periods.

RCP Scenarios	Year	C <sub>4</sub> Grassland	C <sub>3</sub> Grassland	Woodland	Evg. forest	Dec. forest	Shrubland	C <sub>4</sub> Savanna	C <sub>3</sub> Savanna
RCP4.5 + fCO <sub>2</sub>	2010s	0.9	1.5	30.4	189.7	142.1	4.0	35.5	36.8
	2050s	0.9	1.8	29.2	191.0	144.0	3.6	38.0	44.7
	2090s	0.9	2.1	24.5	188.1	148.4	3.3	32.6	31.8
	$\Delta$ 2050s-2010s	-1.1	19.5	-4.0	0.7	1.3	-10.9	6.8	21.4
	$\Delta$ 2090s-2010s	4.4	35.1	-19.4	-0.9	4.4	-17.8	-8.2	-13.7
RCP4.5 + eCO <sub>2</sub>	2010s	0.9	1.4	30.7	189.2	142.5	4.0	35.9	37.3
	2050s	1.0	1.5	34.7	204.6	162.9	4.3	48.1	53.2
	2090s	1.0	1.6	29.3	196.4	164.9	4.1	43.2	51.8
	$\Delta$ 2050s-2010s	17.2	5.6	13.0	8.1	14.4	6.0	34.0	42.7
	$\Delta$ 2090s-2010s	12.6	8.3	-4.6	3.8	15.7	2.5	20.4	39.1
RCP8.5 + fCO <sub>2</sub>	2010s	0.9	1.5	30.7	191.1	146.3	3.9	35.8	34.9
	2050s	0.7	1.6	23.5	182.1	134.7	3.3	31.2	28.0
	2090s	0.8	1.6	18.7	175.7	136.4	3.1	28.5	33.2
	$\Delta$ 2050s-2010s	-19.1	4.7	-23.4	-4.7	-7.9	-15.3	-12.8	-19.7
	$\Delta$ 2090s-2010s	-14.6	4.7	-39.0	-8.0	-6.8	-20.0	-20.5	-4.9
RCP8.5 + eCO <sub>2</sub>	2010s	0.9	1.3	31.2	188.3	146.1	4.1	36.5	32.0
	2050s	1.0	1.4	32.1	206.3	162.7	4.0	45.1	47.2
	2090s	0.7	1.1	30.8	206.0	183.4	4.7	49.8	50.7
	$\Delta$ 2050s-2010s	9.9	8.7	2.8	9.6	11.3	-1.5	23.6	47.4
	$\Delta$ 2090s-2010s	-22.0	-12.7	-1.6	9.4	25.6	15.5	36.6	58.2

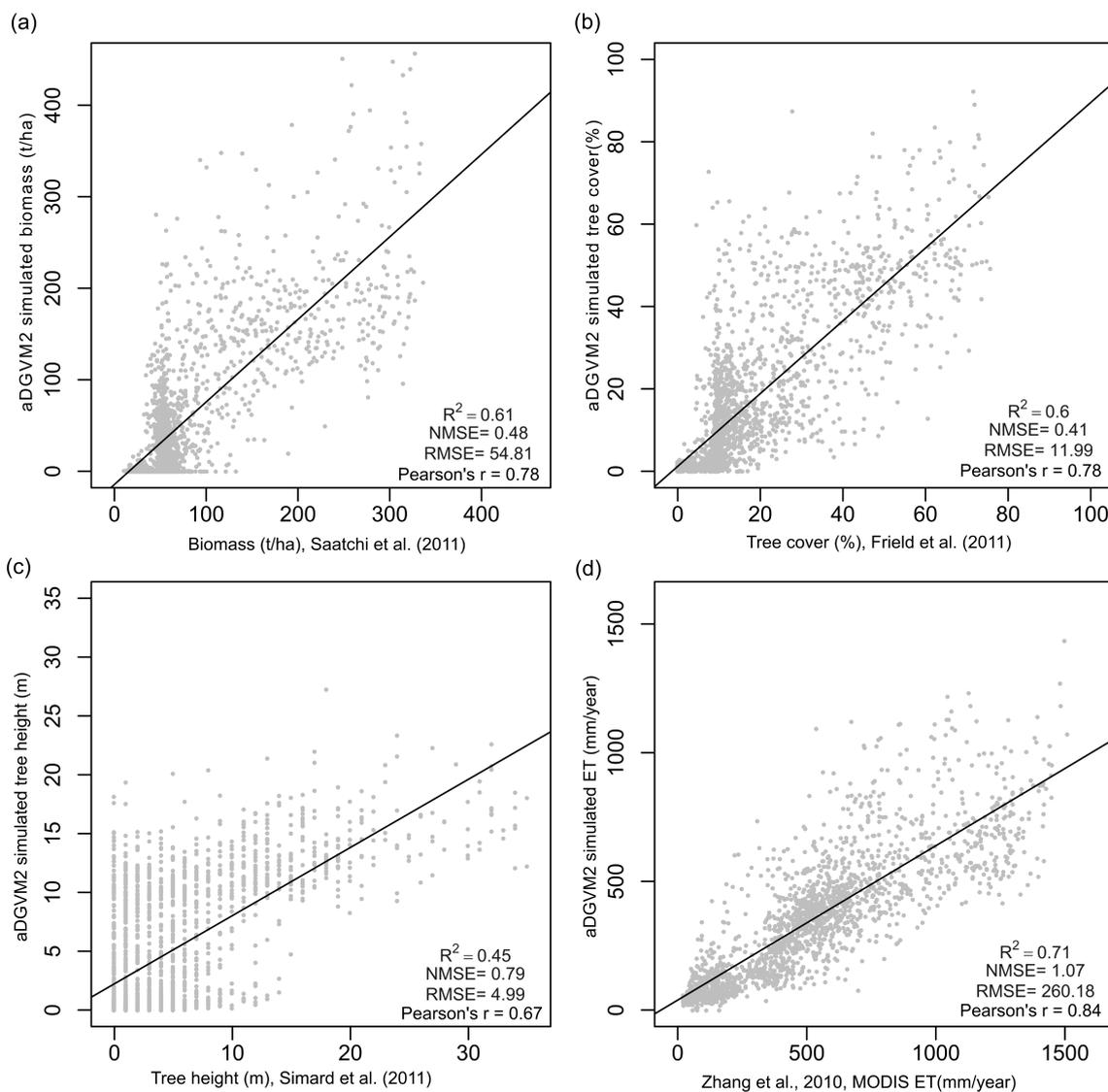


**Table 3.** Biome-level ET normalized to biomass (Ebiomes, mm/kg/year) for the 2000s, 2050s and 2090s, and percent (%) change in Ebiomes from the 2000s to 2050s and the 2000s to 2090s under RCP4.5 and RCP8.5 with fixed and elevated CO<sub>2</sub>.  $\Delta$  indicates percentual ET changes between time periods.

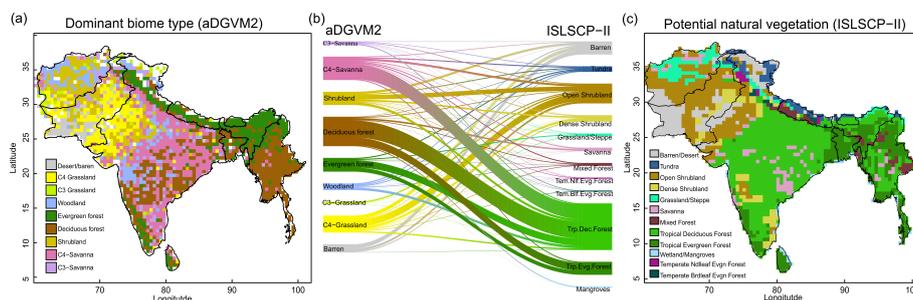
RCP Scenarios	Year	C <sub>4</sub> Grassland	C <sub>3</sub> Grassland	Woodland	Evg. forest	Dec. forest	Shrubland	C <sub>4</sub> Savanna	C <sub>3</sub> Savanna
RCP4.5 + fCO <sub>2</sub>	2010s	186.7	95.5	257	159.7	288.5	183.3	252.5	194.2
	2050s	170.9	80.5	217	157.4	291.3	187.2	244.6	151.9
	2090s	185	72.3	209.6	140.7	289.3	205.2	247.1	179.1
	$\Delta$ 2050s-2010s	-8.5	-15.7	-15.6	-1.4	1	2.1	-3.1	-21.8
	$\Delta$ 2090s-2010s	-0.9	-24.3	-18.5	-11.9	0.3	11.9	-2.1	-7.8
RCP4.5 + eCO <sub>2</sub>	2010s	185.4	93.4	259.7	159.7	288.1	190.9	251.6	188.4
	2050s	161.2	79.7	217	147.8	283	183.2	238.2	153.4
	2090s	164.1	73.4	210.2	138.7	280.4	197.2	236.6	157.1
	$\Delta$ 2050s-2010s	-13.1	-14.6	-16.5	-7.4	-1.8	-4.1	-5.3	-18.6
	$\Delta$ 2090s-2010s	-11.5	-21.4	-19.1	-13.2	-2.7	3.3	-6	-16.6
RCP8.5 + fCO <sub>2</sub>	2010s	172.8	87.4	257.5	160.9	286.5	185.5	244.7	188.1
	2050s	153.7	72.7	243.2	158.3	298.5	195.1	241	162.7
	2090s	195.6	67.6	231.1	162.7	301.3	216	267.5	150.2
	$\Delta$ 2050s-2010s	-11.1	-16.8	-5.5	-1.6	4.2	5.2	-1.5	-13.5
	$\Delta$ 2090s-2010s	13.2	-22.6	-10.2	1.1	5.2	16.4	9.3	-20.1
RCP8.5 + eCO <sub>2</sub>	2010s	177.5	91.1	256.4	162.7	284.5	192.5	243.7	191.7
	2050s	143.9	76.9	235.6	149.4	285.4	184.6	228.8	153.1
	2090s	141.4	59.2	218.3	143.9	284.9	186	242.3	143.2
	$\Delta$ 2050s-2010s	-18.9	-15.6	-8.1	-8.1	0.3	-4.1	-6.1	-20.1
	$\Delta$ 2090s-2010s	-20.3	-35.1	-14.9	-11.6	0.1	-3.4	-0.6	-25.3



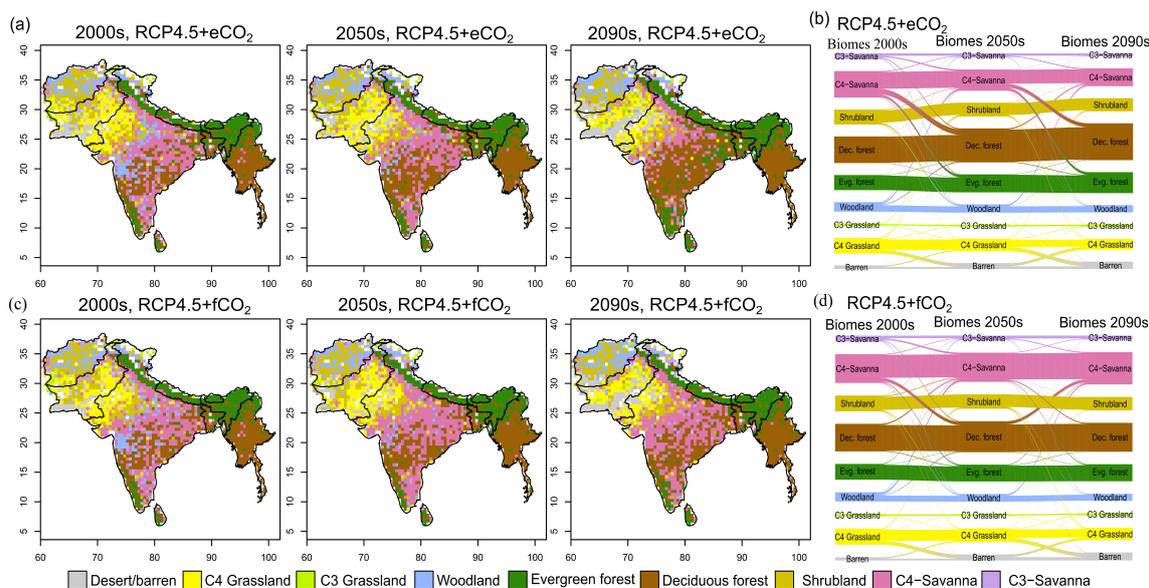
**Figure 1.** : Comparison between aDGVM2 results and data for (a) simulated biomass and Saatchi et al. (2011) biomass and their difference, (b) simulated tree height and Simard et al. (2011) and their difference, (c) simulated tree cover and Friedl et al. (2011) tree cover and their difference and (d) simulated evapotranspiration and Zang et al. (2010) evapotranspiration and their difference. In the figure the first column shows the remote sensing products, the second column shows aDGVM2 results, and the third column shows the difference between simulation and data. For results with masked land use cover see supplementary Figure S2.



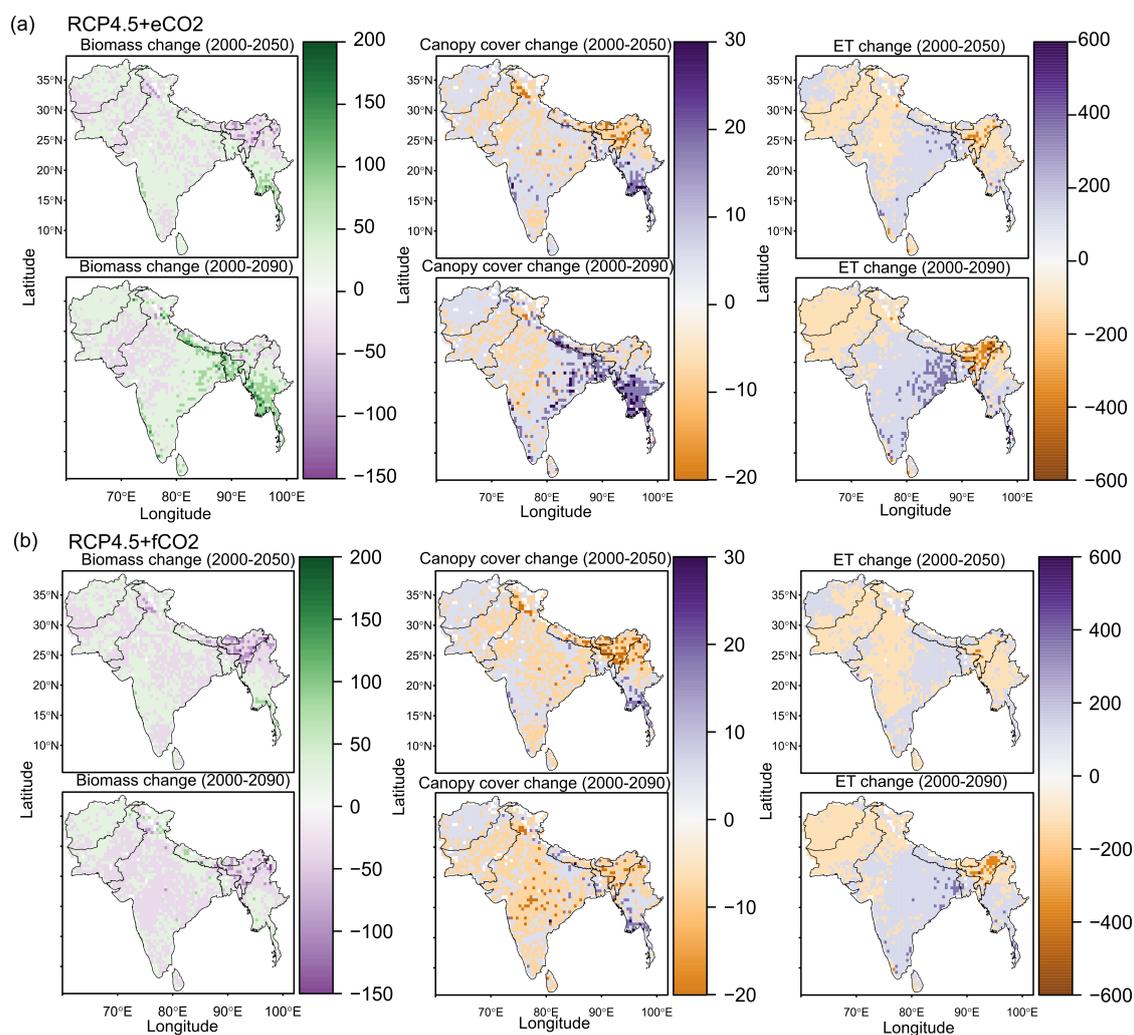
**Figure 2.** Scatterplots for simulated state variables against benchmarking data. (a) Simulated biomass and Saatchi et al. (2011) biomass, (b) simulated tree cover and Friedl et al. (2010) tree cover, (c) simulated tree height and Simard et al. (2011) tree height, and (d) simulated evapotranspiration and MODIS ET (Zhang et al., 2010). NMSE and RMSE are normalized mean square error and root mean square error, respectively. Each point represents one grid cell in the study region.



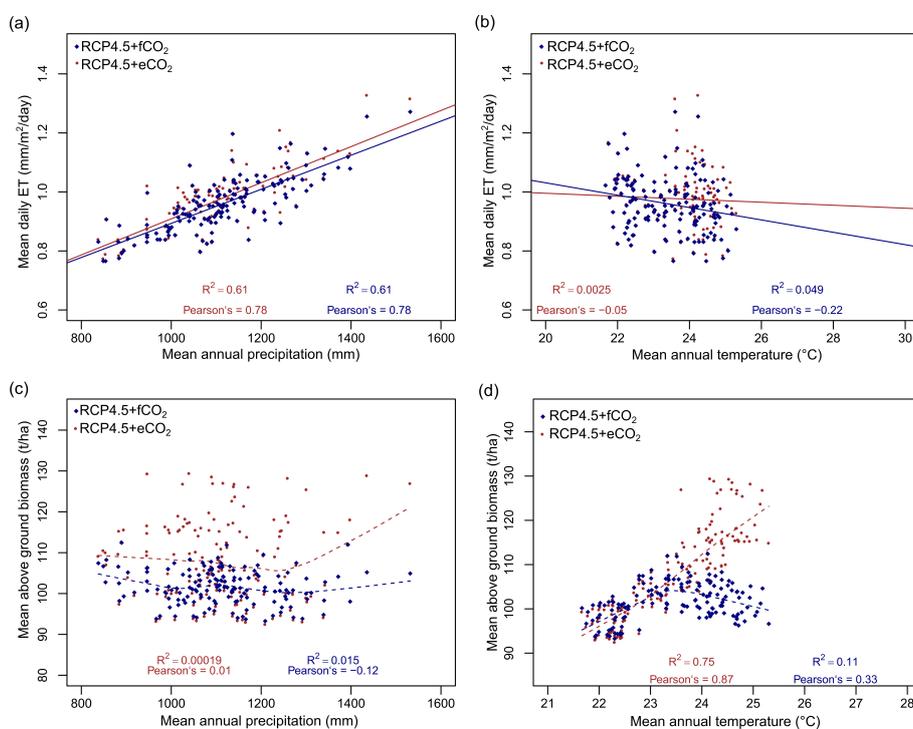
**Figure 3.** Comparison between simulated and observed biome patterns. (a) Simulated dominant biome type, (b) Sankey diagram showing overlap between simulated biomes and potential natural vegetation cover (ISLSCP-II, Ramankutty et al., 2010) and (c) potential natural vegetation. The Sankey graph shows how aDGVM2 biomes and PNV classes overlap.



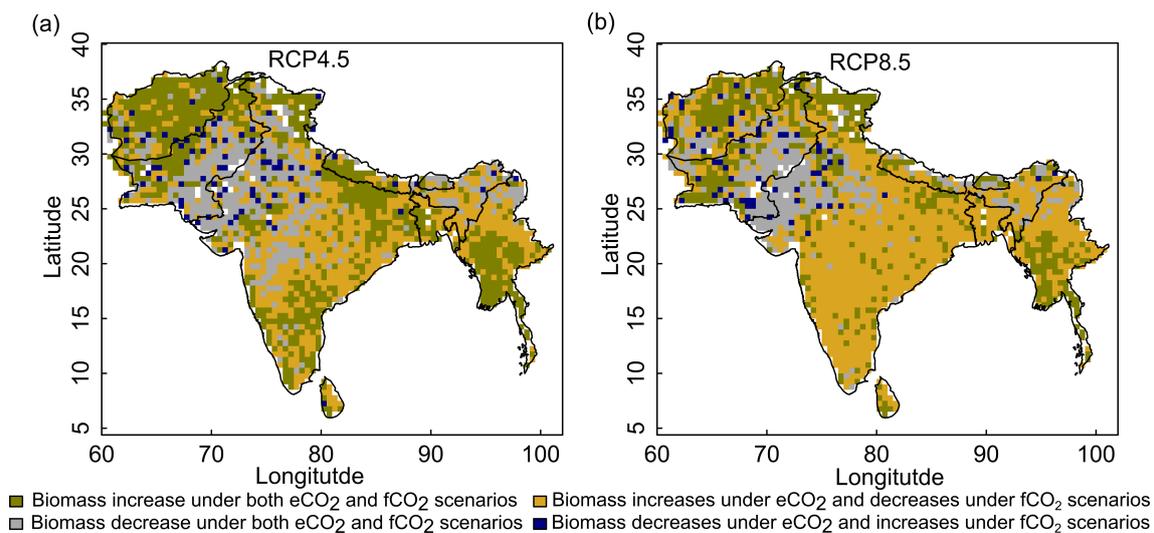
**Figure 4.** Simulated biome distribution for the 2000s, 2050s and 2090s under (a) RCP4.5+eCO<sub>2</sub> and (c) RCP4.5+fCO<sub>2</sub>, and Sankey diagrams showing the fractional cover of biomes and transitions between biomes from the 2000s to the 2050s and the 2050s to the 2090s under (b) RCP4.5+eCO<sub>2</sub> and (d) RCP4.5+fCO<sub>2</sub>. See Figure S6 for simulated biome distribution under RCP8.5.



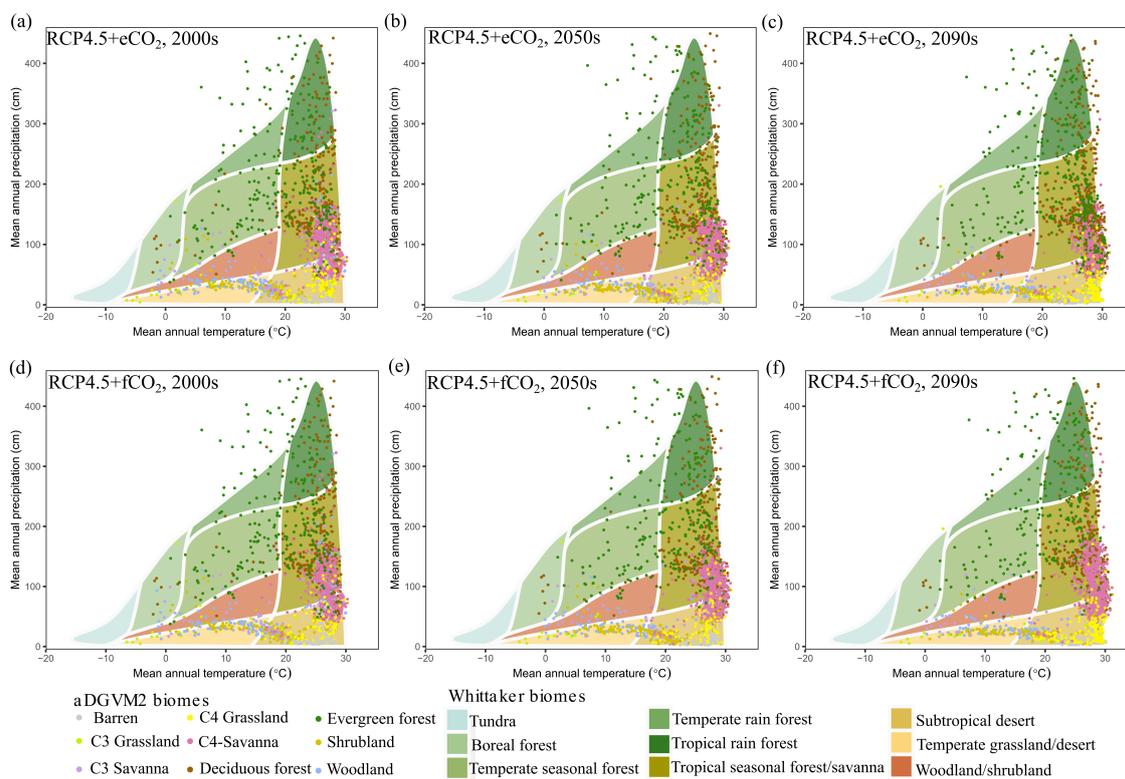
**Figure 5.** Projected change in biomass (t/ha), canopy area (%) and ET (mm/year) between the 2000s and 2050s, and between the 2000s and 2090s under (a) RCP4.5+eCO<sub>2</sub> and (b) RCP4.5+fCO<sub>2</sub>. See Figure S7 for projected change of these variables under RCP8.5.



**Figure 6.** Relationship between (a) evapotranspiration (ET) and mean annual precipitation (MAP), (b) ET and mean annual temperature (MAT), (c) mean above ground biomass and MAP and (d) mean above ground biomass and MAT under RCP4.5. The lines (both solid and dotted) in all figures represent the best-fit regression line. Data points represent spatially averaged ET (a, b) and biomass (c, d) for each year from 1950 to 2099. See Figure S8 for these relation between ET, biomass, MAP and MAT under RCP8.5.



**Figure 7.** Maps showing areas where CO<sub>2</sub>-fertilization compensates biomass dieback caused by climate change between 2000s and 2090s under (a) RCP4.5 and (b) RCP8.5.



**Figure 8.** Simulated climate niches of biomes for the (a) 2000s, (b) 2050s and (c) 2090s under RCP4.5+eCO<sub>2</sub> and (d) 2000s, (e) 2050s and (f) 2090s under RCP4.5+fCO<sub>2</sub>. The simulated biomes are overlaid on the climate envelopes of Whittaker's biomes and are plotted following Ricklefs (2008) and Whittaker (1975). See Figure S9 for projected change in climatic niches of biomes under RCP8.5.